



Evaluating watershed health in Costa Rican national parks and protected areas

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Abstract

Although the biodiversity of Costa Rican national parks, forests, and wetlands has been extensively surveyed, there has not been a watershed assessment that reflects their baseline water quality. Undoubtedly, the influx of 3.1 million visitors annually can lead to deterioration. Additionally, the country's movement toward 100% carbon neutral energy means reliance on capturing water for the production of hydroelectricity. The missing hydrologic data set is of immediate concern, as watershed health predicts total ecosystem health. This field-based project measures eight parameters (pH, temperature, fecal coliform, dissolved oxygen, biochemical oxygen demand, nitrates, total phosphates, turbidity) needed to assign a watershed quality index (WQI) value at nine national parks or protected areas. Overall, the WQI for the systems surveyed reflect good water quality. Results compared with US Environmental Protection Agency (EPA) and World Health Organization (WHO) drinking water standards indicate limited levels of contamination at most sites, with elevated signatures of nitrates, phosphates, turbidity, and/or fecal coliforms at few. The parks selected include coastal lowlands and central highlands; they also experience diverse tourist activities, degrees of use, and forest type that are challenges when managing land sustainably.

Keywords Cost Rica · Hydroelectricity · National parks · Water quality index

Introduction

Costa Rica (51,100 km²) is an equatorial country that is located in Central America (Fig. 1). It is known best for its rich biodiversity and assorted natural landscape. The 500,000 species that are found in this small country represent 4% of the total species estimated worldwide (DeClerck et al. 2010; Kohlmann et al. 2010). The Cordillera de Talamanca mountain range (peak elevation 3820 masl) acts as a drainage divide for all river systems, whereby the eastern slope runoff enters the Caribbean Sea, while the western slope runoff enters into the Pacific Ocean. This extensive

northwest–southeastern mountain belt significantly impacts area climate with the Caribbean flank receiving > 300 cm of annual precipitation and the Pacific flank receiving < 300 cm (Calder 2007).

Approximately, 24% of the land in Costa Rica is designated as a national park or protected area, most of which was gazetted in the 1980s as tourism began to boom. In 2019, tourism in Costa Rica accounted for an estimated total earnings of \$2.8 billion USD that was spent by 3.1 million visitors (per 2019 Census). Of those 3.1 million visitors, approximately half visited national parks and protected areas (Hunt and Harbor 2019; Farrell and Marion 2001). While the promotion of eco-tourism has supplied locals with employment, the country has met with challenges in their sewage infrastructure, the need for more energy, and necessity of water allocation during sustained drought, such as that of the recent day (2014–2017). With that stated, the country has also recognized that an influx of visitors over half of the country's population can cause the landscape to deteriorate if not protected (World Bank 2019).

In 1995, a governmental body was created to help manage the land called the Ministry of the Environment and Energy (MINAE). Subsequently, another group was formed called

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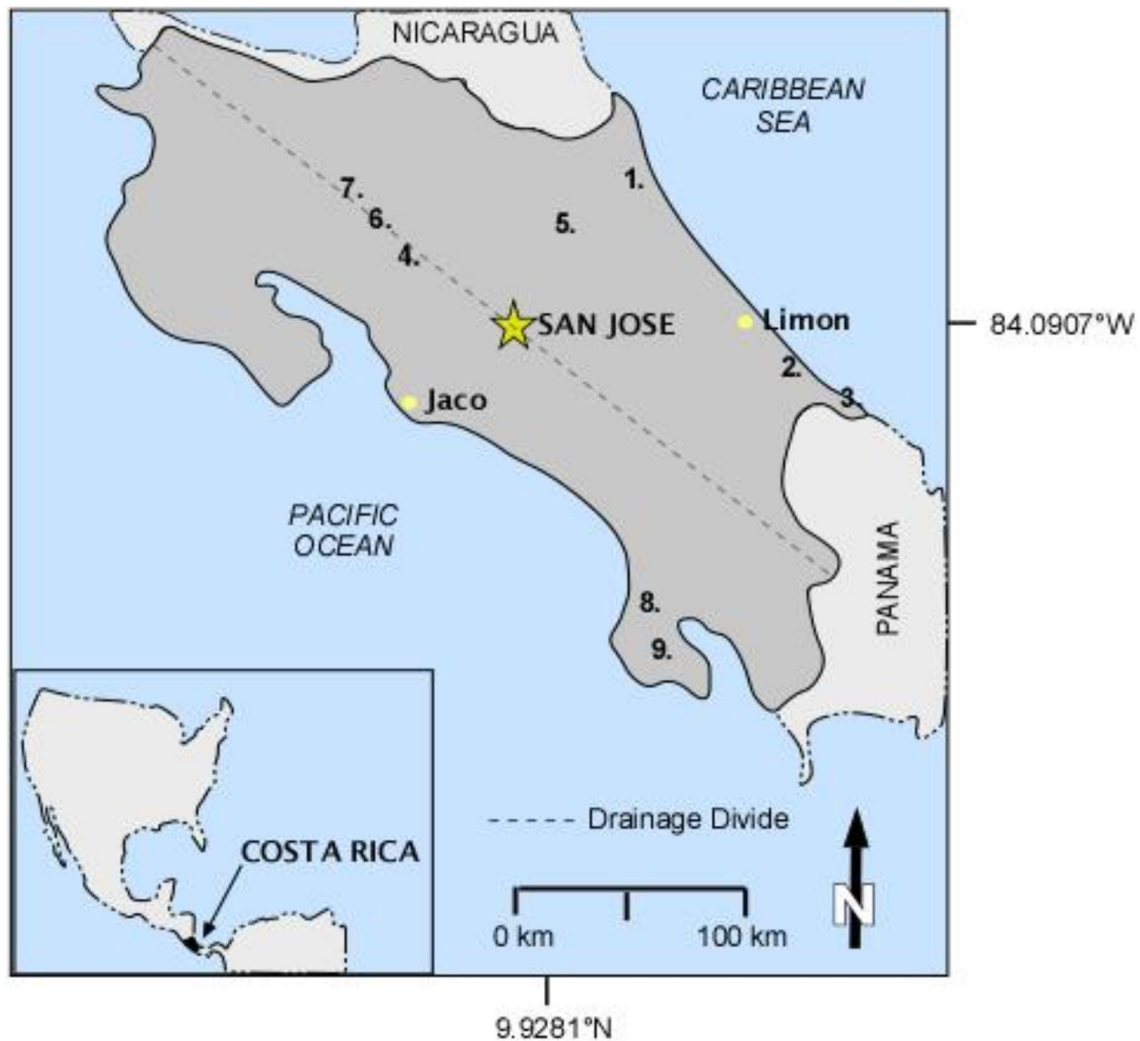


Fig. 1 Location map of sampled national parks (NP), national wetlands (NW), forest reserves (FR), and wildlife refuges (WR). 1. Tortuguero NP, 2. Cahuita NP, 3. Gandoca–Manzanillo WR, 4. Mon-

teverde Cloud FR, 5. Braulio Carrillo NP, 6. Arenal Volcano NP, 7. Tenorio Volcano NP, 8. Terraba Sierpe NW, 9. Corcovado NP

the National System of Conservation Systems (SINAC) which is responsible for the conservation and sustainability use of the country's biodiversity. In total, however, there are over 20 governmental agencies dealing with water resources (Bower 2014). The major dilemma with governance is the absence of a single institution with full responsibility for planning and management of watershed resources and current water laws that have become obsolete (Esquivel-Hernández et al 2018; Guzman-Arias and Calvo-Alvarado 2013).

While Costa Rica has been recognized as one of the more environmentally conscious countries in the developing world (winner of the Ethical Traveler Destination award in 2011 and 2012, for example), the government's push for it to become the first carbon-neutral country by 2021 means increasing hydroelectricity by 20% in the next

5 years (Worldwatch Institute 2017). It is known that dam construction is one of the primary drivers that alters environmental conditions of river ecosystems and may affect its water quality and water self-purification capacity (Brewitt and Colwin 2019; Wei et al. 2009). Water self-purification capacity refers to the maximal capacity of a river segment to take in pollutant without any risk and depends on several factors such as flow volume, flow velocity, pollutant concentration, pollutant decomposition rate, etc. These factors are altered with dam construction. Moreover, Winton et al. (2019) suggest that changes to the stream channel morphology and/or the building of a reservoir can negatively impact water quality through two main processes: the trapping of nutrients and sediments, and thermal stratification. The ecological consequences of dams stem primarily from river fragmentation, stream de-watering, and downstream

alterations (Gierszewski et al. 2020; Anderson et al. 2006; Dynesis and Nilsson 1994). It is also well documented that dams affect the distribution and abundance of aquatic biota, especially migratory species (Jones and Bull 2020; Pereira et al. 2020). Additionally, an alarming and highly referenced climate model created by the University of Massachusetts-Climate System Research Center predicts that the Costa Rican high elevation Pacific slopes and the Caribbean lowlands are likely to receive 30% less precipitation within the next 50–100 years (Stan et al. 2020; Castillo et al. 2017; Karmalkar et al. 2008). Thus, the predetermined functionality of the dams is likely to change within a half of century.

While the biodiversity of the national parks and forested protected areas has been extensively and intensely surveyed, there is a recent country-wide interest on the collection of baseline water quality data through organized watershed assessment in Costa Rica (Mena-Rivera et al. 2017, 2018). Although the building of a hydroelectric facility or dam in a national park (or similar) is unlikely, the concern is that streams adjacent to preserved landholdings are often targeted for construction and channelization that alters the flow entering the park (Opperman et al. 2019; Anderson et al. 2006). Moreover, according to Bower (2014), the country has not yet developed a plan of sustainable water use whereby considering its future demands against the availability and quality of water. Unanswered, this could prove problematic as only 8.2% of the raw sewage in Costa Rica is treated before it is released into the environment, highlighting the need for sanitation infrastructure development and monitoring (Mena-Rivera et al. 2018; Bower 2014). Forest cover and wetlands adjacent to biological corridors and national parks continue to be threatened by contaminants: from heavy agricultural activities such as sugarcane, banana and pineapple plantations, rice, citrus, melon, and livestock (Mena-Rivera and Quirós-Vega 2018; Allen and Vásquez 2017; Mitsch et al. 2008). In a 2013 ecosystems review, Costa Rican hydrology experts, Guzman-Arias and Calvo-Alvarado, suggest that data collection on the watersheds is the missing link to manageable and sustainable use. Additionally, data sets representing water quality is an immediate concern because watershed health acts as a predictor for total ecosystem health (Kolok et al. 2020; EPA 2016; Woodall et al. 2011). Vibrant watersheds provide critical services, such as clean drinking water, productive fisheries, and outdoor recreation, which support economies, environment, and quality of life. Understanding that the quality of the water is the ‘canary in the cage’ has been the message overlooked by undeveloped and developing countries (Jovanelly et al. 2015).

Overall, the aim of this project is to highlight the importance of including hydrological system health in Costa Rican forest conservation. The project is timely because the water quality index (WQI) has not been used to communicate

watershed health in Costa Rica. The WQI is a unitless number ranging from 1 to 100 that reflects the overall health of a system by assigning weighted values to the eight parameters (pH, temperature, fecal coliform, dissolved oxygen, biochemical oxygen demand, nitrates, total phosphates, turbidity) measured using field instrumentation. A higher WQI number is indicative of better water quality (100–90 is deemed excellent water quality, for example). Because WQI combines parameters into a single rank value (poor, moderate, etc.), the WQI serves as a straightforward way of communicating changes in watershed health to forest communities and to measure changes in national parks and protected areas.

Study areas

The national parks and protected areas selected for this study reflect both the Caribbean (Tortuguero National Park, Cahuita National Park, Gandoca-Manzanillo Wildlife Refuge) and Pacific (Corcovado National Park, Térraba Sierpe National Wetland) lowlands, as well as the Central Highlands (Monteverde Cloud Forest Reserve, Arenal Volcano National Park, Braulio Carrillo National Park, Tenorio Volcano National Park) (Fig. 1). These nine national parks or protected areas also demonstrate the diversity of touristic activities, varying degrees of visitation, and assortment of forest types that challenge the Costa Rican government when managing land sustainably. Additionally, these preserves are contained within distinct watersheds (UNESCO 2007).

Caribbean coast

The Costa Rican Caribbean coast is located on the country’s eastern shoreline and is the result of subduction process between the Cocos and Caribbean plates and the development of volcanic arcs, shallow deposits, and turbidity environments around 65 Ma. The study area encompasses two geological regions, the Caribbean Basin and Limón Basin, to form a 50 km-wide depression extending from broad alluvial plains in the north to marine carbonate platforms of the southeast. The carbonate platforms are of marine origin, formed by Plio-Pleistocene reef limestone rocks that have been gradually raised by slow uplift punctuated by sudden vertical movements due to regional tectonic activity (Quezada-Román and Pérez-Briceno 2019). The most common soil orders found along the coast are of low-porosity types such as entisols and inceptisols, although organic-rich histosols have been identified in some flood plain areas. The tropical forests (average rainfall 477 cm) along the Caribbean coast provide an environment to support extensive pastures for livestock and banana plantations. The coastal lagoons of the Caribbean coast exhibit characteristics of estuaries that

have been closed by sedimentation, like that at Tortuguero National Park (Location 1; 10°26'40" N, 83°30'36" W). This national park has over 32 km of coastline that is a habitat for sea turtles. Tortuguero National Park (311 km²) is bordered by the Barra del Colorado Wildlife Refuge to the north and the Parismina River and Cariari National Wetlands to the south. The Tortuguero River is a primary freshwater source for this area and flows into the Caribbean Sea. The morphology of Location 2 (Cahuita National Park; 9°44'11" N, 82°50'19" W) and Location 3 (Gandoca-Manzanillo Wildlife Refuge; 9°37'48" N, 82°38'56" W) is best described by their white sand beaches now in place due to the coastal erosion of coral reefs. Cahuita National Park is 11 km², but is most known for its beach access for offshore protected coral reefs. The remote Gandoca-Manzanillo Wildlife Refuge protects 15 km of the southern Caribbean coast, extending from Manzanillo to the Panama border. The refuge encompasses 50 km² of lowland tropical rainforest, wetlands, and mangrove swamps.

Central highlands

The Cordillera de Talamanca extends 200 km along the Central Highlands of Costa Rica and is the tallest mountain range in Costa Rica with volcanic peaks reaching 3500 m (average 2000 m). The complicated formation of the range introduces Eocene age (~33.9 Ma) volcanic rocks that are followed by igneous intrusive batholiths and later the erosional formation of rocks with sedimentary lithologies (Alvarado and Gans 2012). The shape of drainage basins within the Cordillera de Talamanca were further influenced by Pleistocene glacial and periglacial landscapes that scoured deep U-shaped valleys that are sometimes lined with an impermeable loess layer. The overall relief of the mountain range is unusual in that the shape is asymmetrical in cross section: the SW flank is steep, and the NE slope is moderately inclined.

A classic mountain cloud forest on the southern side of Cordillera de Talamanca called Monteverde Cloud Forest Reserve (Location 4: 10°18'08" N, 84°47'45" W) was sampled in both 2018 and 2019. Monteverde Cloud Forest Reserve was established in 1972 with an area of 3.3 km² and has expanded to cover 142 km² at the present day. This area contains eight different ecotones atop of the Talamanca Continental Divide that separates the drainage from the Pacific Ocean on the western flank from that of the Caribbean Sea on the east. This site receives 260 cm rainfall annually and is at an elevation of ~1600 masl (Guswa et al. 2007).

Three volcanic national parks were sampled that include: Arenal Volcano National Park (location 5: 10°31'59" N, 84°43'13" W), Braulio Carrillo National Park (location 6: 10°09'06" N, 84°05'39" W), and Tenorio Volcano National Park (location 7: 10°42'43" N, 85°01'04" W). Arenal

Volcano National Park is 120 km² and is located within the larger 2039 km² Arenal Conservation Area. Arenal Volcano is an active andesitic stratovolcano that reaches 1140 m. The volcano was dormant for hundreds of years prior to its last eruption in 1968. Fumaroles and hot springs are still present in the area. Nearby, Lake Arenal is the largest lake in Costa Rica at 85 km² with a depth between 30 and 60 m. In 1979, this reservoir tripled in size when it was expanded to accommodate a hydroelectric dam that was installed. The Lake Arenal Dam produces approximately 12% of the country's total electricity and is the third largest producer in the country at 180 MW annually. The area receives approximately 406 cm per year.

Braulio Carrillo National Park was gazette in 1978 and covers 183 km², including the inactive volcano that peaks at 2900 m. The forest is supported by 203 cm per year of rainfall and is unique in that the habitats represent a cloud forest, a regenerating tropical rainforest lowland and a tropical rainforest upland. The Sucio River is the main river that runs through the park and is named after sulfur deposits found on the nearby Volcano Irazú. The Sucio River is a tributary of the much larger Río San Juan.

Tenorio Volcano National Park was created to protect the pre-montane forest from logging in 1976 and is 130 km². The park contains four volcanic cones with a maximum height of 1916 m. The volcanic complex has been dormant since 1997 when microseismicity was last reported. Running through the park is a popular tourist destination called Río Celeste that is supported by the 475 cm of rainfall the area receives annually. It is a 14 km-long river that is formed by the convergence of Quebrada Agria and Río Buena Vista. Río Celeste is a milky blue color from the high amounts of aluminum and silicates that are suspended in the water (Castellón et al. 2013). Additionally, hot springs and fumaroles can be found throughout the national park.

Pacific coast

The Pacific coastal region of Costa Rica is characterized by an active convergent plate margin of complex origin starting in the mid-Miocene approximately 8 Mya. Here, a collisional boundary has formed between the Caribbean and Cocos tectonic plates, whereby the rate of subduction is estimated to be 9.5 cm per year (Wells et al. 1988). The intensity of this subduction zone has translated into significant amounts of Quaternary deformation seen along the southwestern coastal region that include fault scarps displacing fluvial terraces and changes to the deposition of colluvial soils in areas of collisional response (Harpster et al. 1981). These shifts in surficial processes influence drainage basin shape by creating steep slopes and irregular longitudinal river profiles that are characterized by concave and convex reaches.

The Pacific Coast of Costa Rica has a large number of embayments that provide shelter from wind and wave activity that can favor mangrove development (Niekerk et al. 2020; Polanía 1993). Uniquely, mangroves are more developed in this ecoregion than further up the coastline due to the higher rate of freshwater inflow that reduces salt accumulation in the mangroves by increasing evapotranspiration (Jimenez 1999). Térraba Sierpe National Wetland (Location 8: 8°50'17" N, 83°33'14" W) is the largest mangrove system in Costa Rica (270 km²) and is among the largest in Central America. This National Wetland includes the Térraba-Sierpe river delta that was used for agricultural planting for six decades prior to the wetland being gazetted in 1994 with an annual net loss of 40 ha per year between 1948–1994 (Acuna-Piedra and Quesada-Roman 2017). Rainfall and tidal fluctuations directly impact mangrove ecosystems. This ecoregion receives an average of 368 cm of rain annually and has a mean tidal gain of 3.5 m (Polanía 1993).

The Osa Peninsula is south of the mangrove forest and is home to the largest remaining primary forest (424 km²) on the Pacific Ocean at Corcovado National Park (Location 9: 8°30'22" N, 83°35'24" W). About 85% of this national park is considered pristine and it is bisected by numerous small streams and rivers that drain the ridges along the eastern park border to the Pacific Ocean. Other inland waters include a large marsh, a palm swamp, and several small estuaries (Constantz et al. 1981). This lowland tropical forest (average rainfall 500 cm) became protected in 1975 and was saved from adjacent logging activity.

Materials and methods

Data for eight parameters, dissolved oxygen (DO), biochemical oxygen demand (BOD), pH, turbidity, nitrates, phosphates, fecal coliform, and temperature, were used to calculate the WQI as specified under the National Sanitation Foundation guidelines (Mitchell and Williams 2000; Brown et al. 1970). WQI reflects the overall health of a water system by assigning weighted values to the parameters.

Water quality data were collected in each of the national parks or protected areas by the research team once or twice over two wet seasons in 2018 and 2019 (Fig. 2). During each site visit, the researchers collected water samples. The sampling locations depended upon accessibility. The six water quality parameters (DO, pH, turbidity, nitrates, phosphates, temperature) were measured on site using field instrumentation (Fig. 3). A handheld Texas Instrument Nspire CX calculator was used with corresponding dissolved oxygen, pH, and turbidity probes that connected directly to the instrument allowing for immediate field sampling. All probes were calibrated prior to each field visit using the manufacturer's instructions. A portable LaMotte SMART

3 Colorimeter was used to measure nitrate and phosphorus at the field location. The instrument was calibrated for each parameter independently prior to use according to the instrument guidelines. Nitrates were determined using a zinc reduction method and the phosphorus was established using the vanadomolybdophosphoric acid method. Temperature was measured on-site using a mercury thermometer.

At each sample site, water was collected in 250 ml Nalgene amber sample bottles and transported to the laboratory. The fecal coliform test was completed in the laboratory (usually within 4 h of sample collection) using 3 M® Petrifilm Coliform Count Plates by inoculating 1 ml water sample according to the manufacture's guidelines. The plates were incubated at room temperature for 24 h. After 24 h the colonies on the slide were counted and recorded.

The remaining water samples were capped tightly and placed in a dark cabinet for 5 days and incubated at room temperature. The water sample was tested again after 5 days for DO using the TI-Nspire CX calculator and probe, and BOD calculated. After sampling, a baseline WQI (that represents seasonal variability) for each forest reserve was established.

Data analysis was carried out by computing the descriptive and inferential statistic; the latter was done by a permutation test (Good 2009; Blair et al. 1994) where differences of mean values for the WQI and significant parameters were tested among national parks locations. Also multivariate analysis (principal component analysis, PCA) was computed to identify which parameters can explain the main variability within the obtained results for WQI.

Results

The WQI monthly averages for each of the nine national parks or protected areas are displayed in Fig. 2; monthly averages of the eight WQI parameters measured are found in Fig. 3.

Caribbean locations

The WQI at Tortuguero National Park (Location 1, Fig. 1) ranged between 73.0 and 77.0%. Of the three sample sites, all fell in the good range. Dissolved oxygen ranged between 5.00 and 5.80 mg/L and was lowest in the mangrove swamp (5.30 mg/L). BOD ranged between 0.10 and 0.3 mg/L, ranking the overall BOD levels between 99 and 100%. The water acidity ranged between a pH of 5.98 and 6.26 resulting in WQI values for pH from below average (54%). The temperature of the surface water sampled was 26–28 °C. Nitrates were found in all of the samples at Tortuguero National Park (1.0–6.0 mg/L). Site TG1, a mangrove swamp, had elevated nitrates compared to sites TG2 and TG3. The lowest

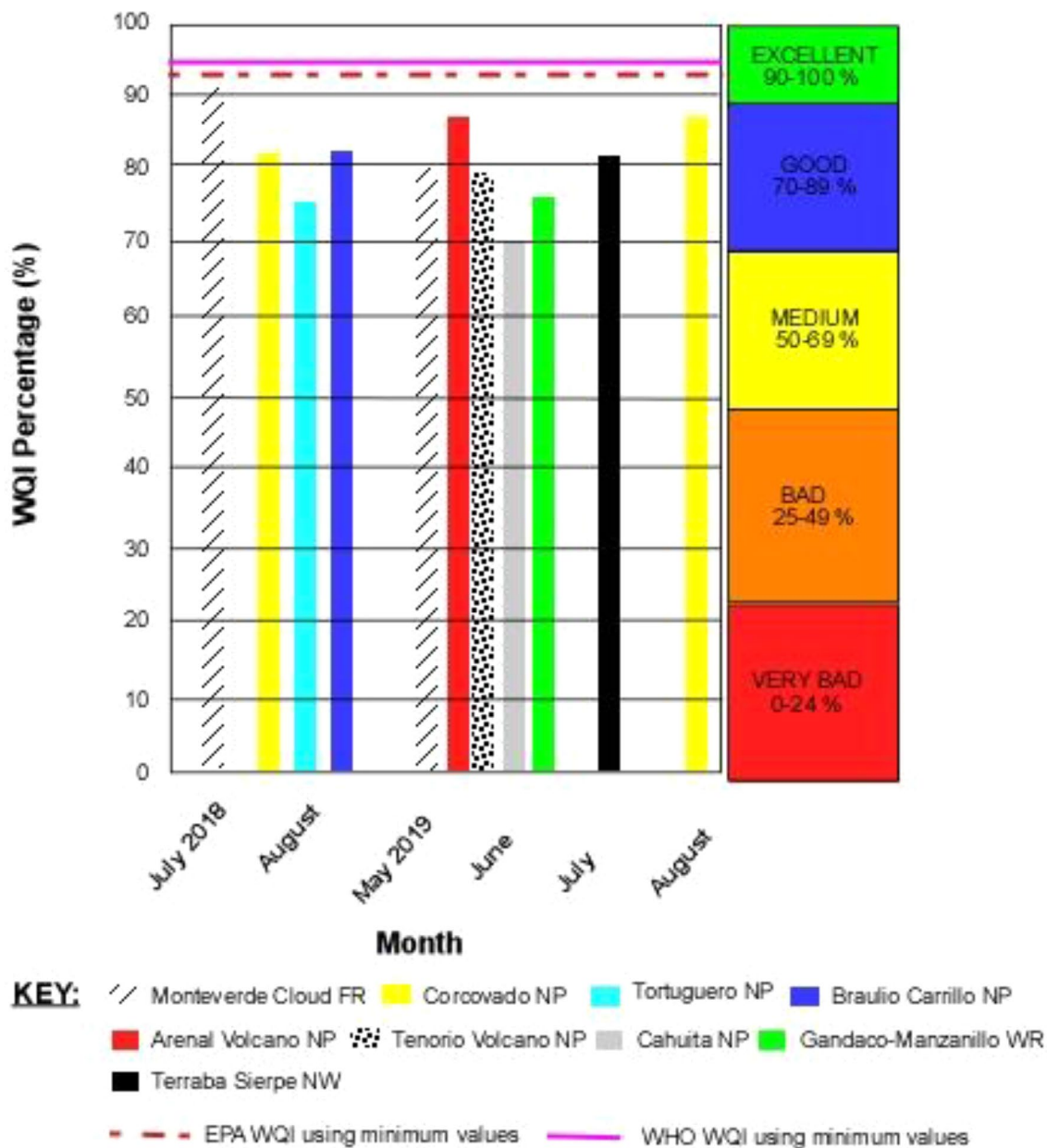


Fig. 2 Average monthly water quality indices for nine national parks (NP), national wetlands (NW), forest reserves (FR), or Wildlife refuges (WR) as compared to the USEPA and WHO WQI calculated using minimum values from Table 1 for drinking water

nitrate (1.0 mg/L) of all the sample locations was found in the Tortuguero River. Phosphates ranged between 1.0 and 3.0 mg/L. July recorded the overall lowest levels of phosphates. Fecal coliform levels at Tortuguero National Park produced excellent WQI index values at 100.0% as all of the samples were devoid of fecal coliform. The water from Tortuguero National Park had turbidity as high as 36.90 NTU and the lowest documented turbidity was 32.90 NTU.

The average WQI of Cahuita National Forest (Location 2, Fig. 1) ranged between 58.0 and 76.0%, representing medium to good water quality. Measured dissolved oxygen

ranged between 2.7 and 8.40 mg/L. Low dissolved oxygen levels occurred at CH3 (2.70 mg/L) and the other six sites averaged 7.8 mg/L. Little change occurred between the initial and final DO in the BOD test; all values reported are < 1.0 mg/L. The pH ranged between 7.09 and 7.58. The highest nitrate concentration was 50.0 mg/L at CH1, while the lowest was 11.0 mg/L at CH6. Phosphate concentrations ranged between 7.0 and 27.0 mg/L. Fecal coliform concentrations were detected at all sites and ranged between 25 and 90 colonies/100 ml. Turbidity was also elevated (17.30–78.90 NTU).

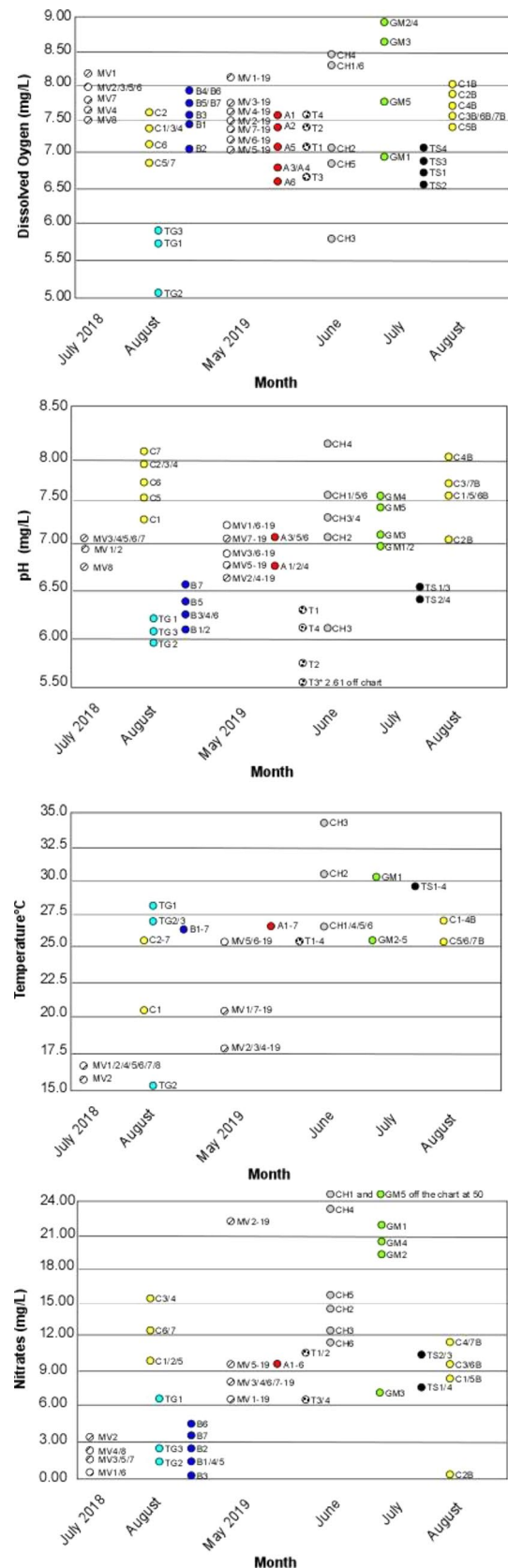
Fig. 3 Values measured for each parameter at field sites correlating to Figs. 1 and 2. Monteverde Cloud FR (MV), Corcovado NP (C- for 2018 data; CB for 2019 data), Tortuguero NP (TG), Braulio Carrillo NP (B), Arenal Volcano NP (A), Tenorio Volcano NP (T), Cahuita NP (CH), Gandoca-Manzanillo WR (GM), Térraba-Sierpe NsW (TS). The values are grouped in cases where the range is 0.5 mg/L or less, 1 NTU or less, 5 colonies/100 mL or less, or 1 °C

The average WQI at Gandoca-Manzanillo Wildlife Refuge (Location 3, Fig. 1) was good (73.0–85.0%). Measured dissolved oxygen ranged from 7.80 to 9.0 mg/L at all sites except GM1, which showed lower dissolved oxygen levels (6.90 mg/L). BOD at Gandoca-Manzanillo Wildlife Refuge ranged between 0.10 and 0.20 mg/L. The pH was within 6.86–7.49. Elevated nitrates (8.0–50.0 mg/L) and phosphates (1.0–9.0 mg/L) were detected at all five sites.

Central highland locations

Sampling at Monteverde Cloud Forest Reserve (Location 4, Fig. 1) occurred twice in the wet season: July 2018 and May 2019. The average WQI at Monteverde Cloud Forest Reserve ranged between 88.0 and 92.0% in 2018 and 74 and 83% in 2019, or good classification. Dissolved oxygen ranged between 7.7 and 8.1 mg/L (2018) and 7.0 and 8.2 mg/L (2019). The dissolved oxygen concentrations were higher at all sample sites in 2018. Microorganism activity was recorded by measuring BOD (0.10–0.20 mg/L) for both years. The values measured for pH were similar at all seven sites when comparing the 2018 data to 2019; the pH ranges from 6.8 to 7.01 in 2018 and 6.78 to 7.09 in 2019. The 2018 data set for nitrates and phosphates reveals much lower levels than reported for 2019 at the same sites. In 2018 the range of nitrates were non-detectable to 3.0 mg/L, compared to 7.0–9.0 mg/L in 2019 with an outlier value of 22.0 mg/L reported at MV2-19. Phosphates showed the same pattern whereby 2018 values were lower (non-detectable to 0.01 mg/L) than those recorded in 2019 (1.0–3.0 mg/L). The turbidity levels tend to overlap from year to year (14.20–32.0 NTU in 2018; 15.0–43.0 NTU in 2019); however, the averaged value of turbidity from all the sites is greater in 2019 (29.3 NTU) than 2018 (19.30 NTU). Additionally, the overall averaged WQI dropped from 90.6% (2018) to 79.8% (2019).

Sampling in Arenal Volcano National Park (Location 5, Fig. 1) occurred on the adjacent surface water reservoir at six sites that were randomly selected. The dissolved oxygen ranged from 6.60 to 7.50 mg/L and the pH ranged from 6.78 to 7.08. The average turbidity was 28.5 NTU spanning from 29.9 to 36.20 NTU. Although phosphates were found to be below 1.0 mg/L at all sample sites, the nitrates were elevated between 7.0 and 8.0 mg/L. Thus, the turbidity and the nitrates were the two elevated variables that brought down the overall WQI to 85.2%.



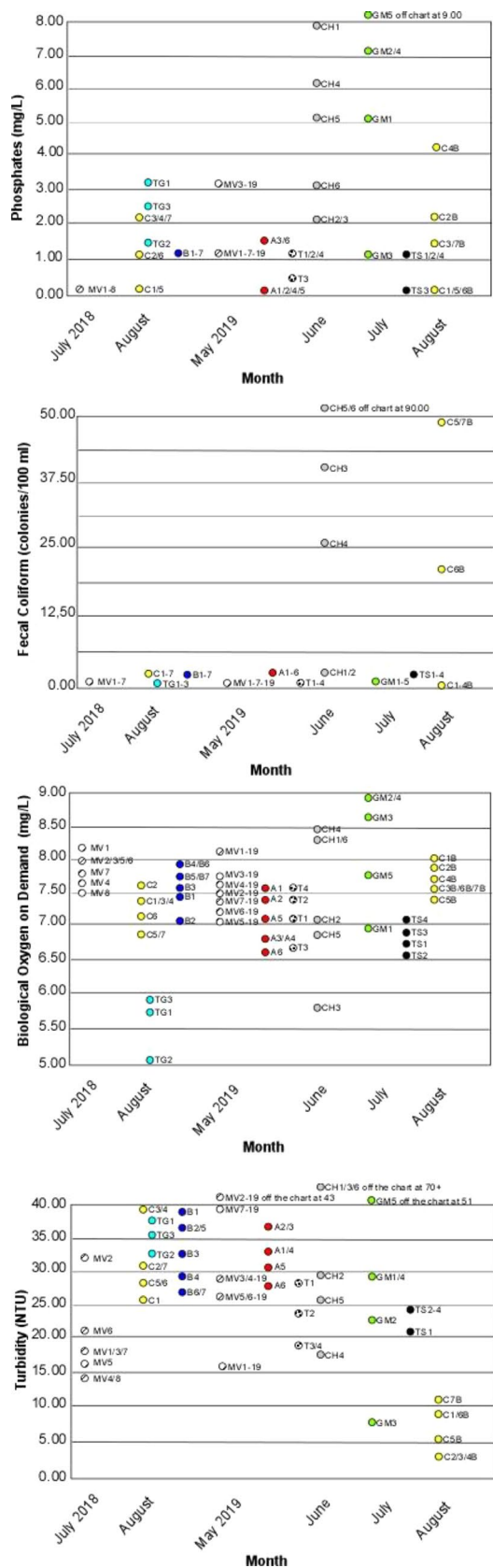


Fig. 3 (continued)

The seven sites sampled at Braulio Carrillo National Park (Location 6, Fig. 1) determined an overall WQI of 81.9%, or better. The pH for all sample sites was slightly acidic ranging from 6.13 to 6.39. The water temperature in all streams sampled was found to be between 22 and 23 °C, while the turbidity was elevated ranging between 28.0 and 38.0 NTU. The highest dissolved oxygen level (7.8 mg/L) was measured at B4 and B6. No fecal coliform levels were reported for this national park.

Of all of the national parks visited for this study, Tenorio Volcano National Park (Location 7, Fig. 1) has a site with the lowest pH (2.61) logged at T3. The other three sites were also slightly acidic (5.74–6.44). Prominent values were established for nitrates that ranged from 6.0 to 9.0 mg/L, while phosphates remained negligible. The silica-rich waters lent themselves to higher turbidity counts that averaged 22.5 NTU and ranged from 19.0 to 27.1 NTU. The overall WQI at Tenorio Volcano National Park is 78.0%.

Pacific locations

Despite having the warmest temperatures recorded (29–30 °C) of all the national park locations, Térraba Sierpe National Wetland (Location 8, Fig. 1) had an overall average WQI that was good (80.8%). The pH tended to be slightly acidic (6.40–6.50) and the dissolved oxygen low (6.5–7.0 mg/L). While the phosphates were 1.0 mg/L or less, the nitrates ranged from 8.0 to 11.0 mg/L.

Seven sample sites were visited at Corcovado National Park (Location 9, Fig. 1) over two field seasons at two different park access points. In 2018 we accessed the southernmost park entrance from HWY 245 and in 2019 our access point was from Sirena Biological Field Station. The overall average WQI in 2018 was 81.0% and in 2019 was 84.3%. The lower WQI values in 2018 are impacted by the lower dissolved oxygen values (6.7–7.6), and both elevated nitrates (10.0–15.0 mg/L) and turbidity (25.00–40.90 NTU) when compared to the 2019 data. Despite the DO levels being within a normal range (7.4–7.8 mg/L) at all seven sample sites in 2019, there were slightly elevated levels of phosphates (0.0–4.0 mg/L), nitrates (1.0–11.0 mg/L), and turbidity (3.0–10.0 mg/L). No fecal coliform was found at sites C1B, C2B, C3B, and C4B; however high levels were found at C5B (50.00 colonies/100 ml), C6B (20.00 colonies/100 ml), and C7B (50.00 colonies/100 mL). No fecal coliform was found at any of the sites in 2018.

Multivariate and inferential analysis

PCA analysis shows that there are three significant dimensions (Dim) within the data (eigenvalues higher than 1.0; Yeomans and Golder 1982), accounting for the 64.9% of the cumulative variance percentage (Fig. 4a). Among the

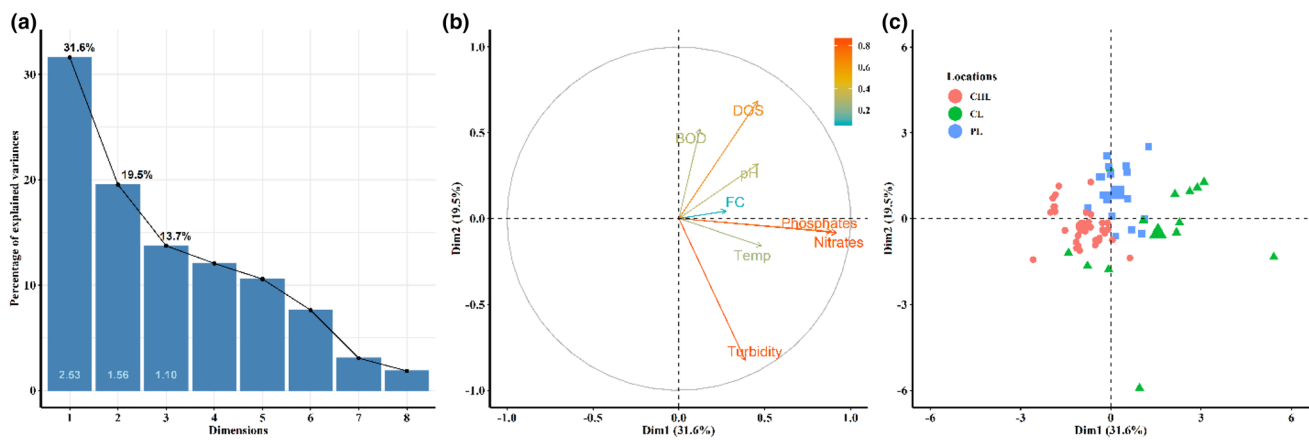


Fig. 4 Principal component analysis comparing **a** cumulative variance percentage, **b** phosphates and nitrates, and **c** location

two first Dim (Fig. 4b), phosphates and nitrates have the highest loading factors (> 0.75 ; Liu et al., 2003) in the Dim1 (0.89 and 0.92, respectively), explaining 31.6% of the data variance; in Dim 2, dissolved oxygen saturation (DOS) and turbidity are the parameters with the highest loading factor (0.68 and -0.89), and Dim 3 accounts for the 13.7% of variance. Nevertheless, only fecal coliform has a high factor loading (0.87).

Spatial variability among the national park locations is explained in Fig. 4c. Caribbean locations have more variation within Dim 1; also, it is noticeable that they are well separated from the Pacific coast and Central Highlands, in accordance with major loading factor in phosphates and nitrates. Analysis in Dim2 shows that there are not clearly separations among the locations; however, the majority of points for the Caribbean locations have a negative factor loading in this dimension, relating with the turbidity contributions.

PCA shows that a significant proportion of the total data variance could be explained among the five parameters that had been measured, all of them being related with agricultural activities and/or the lack of a good wastewater sanitation infrastructure (Barrenha et al. 2018; Huang et al. 2017). Phosphates and nitrates are the significant parameters that influence the WQI variability, since these had major contributions in factor loading terms for Dim1. Similarly, DOS and turbidity have significant contribution; however, it is important to stress that positive and negative factor loadings are related with the opposite trend for the variables, which means that increases of turbidity might be linked with decreases of DOS. The latter could be caused by the organic matter that is suspended in high turbidity conditions (Mena-Rivera et al. 2017).

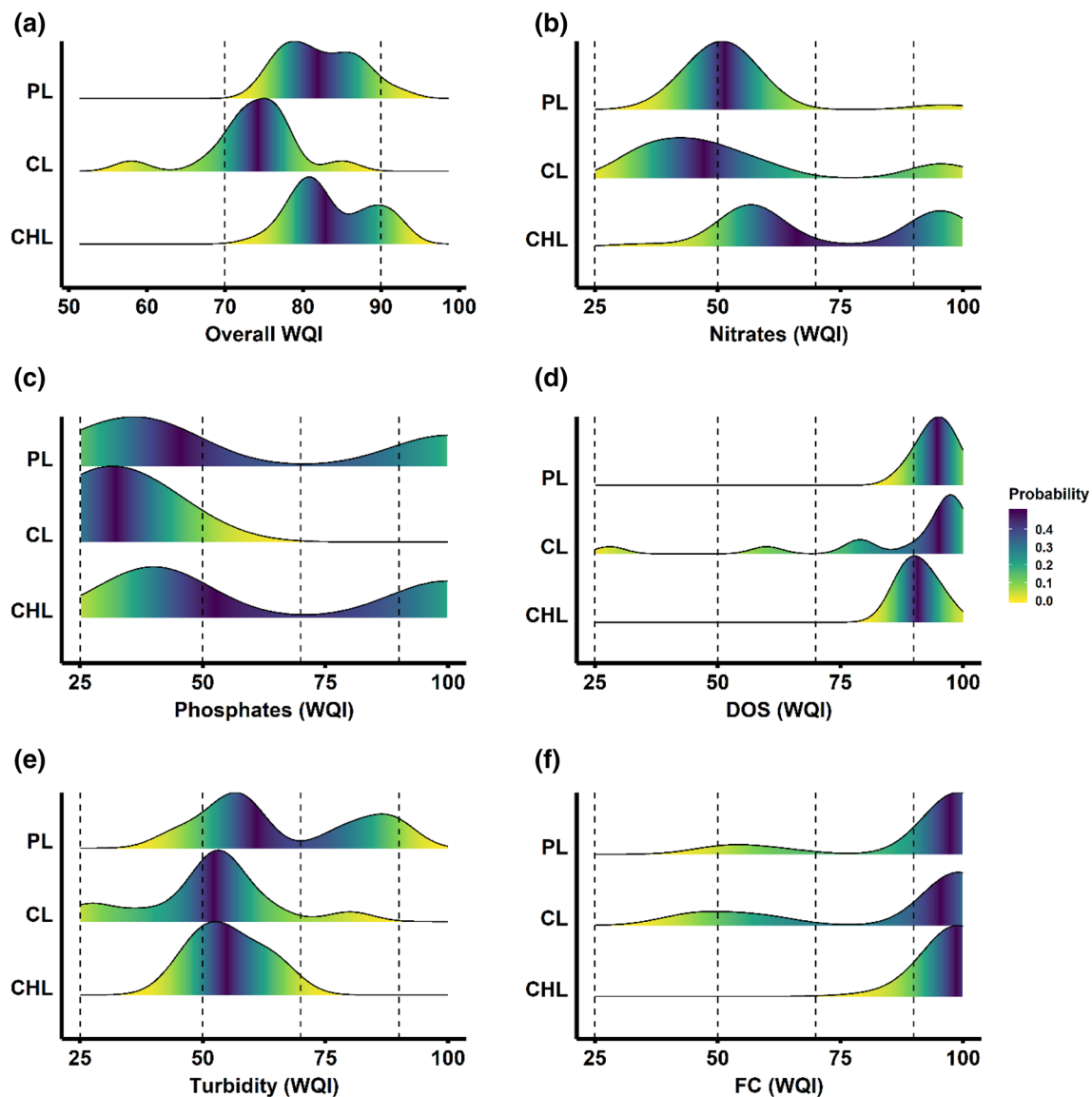
An evaluation through permutation inferential statistic test was done for the results that were obtained for the WQI as an overall and for the parameters that resulted with the

highest loading factors in the significant dimensions of the PCA, revealing that significant difference was found in the majority of those parameters.

Figure 5a shows the distribution for the overall WQI scores. Caribbean locations have the lowest mean and it presents significant differences between the Pacific Coast and Central Highlands ($F = 19.30$, $p < 0.05$). Figure 5b, c shows the distribution of the WQI scores for nitrates and phosphates. In both cases, CL has significant lower mean values (47 and 19, respectively) than the other locations ($F > 11$, $p < 0.05$). However, it is important to address that the mean distribution for nitrates did not differ for CL and PL (diff < 6.2 , $p > 0.05$), the latter suggesting that nitrate input rate and sources in both locations could be similar. DO and turbidity distributions are shown in Fig. 5d, e. Non-significant differences were found for the DO mean WQI score ($F = 2.4$, $p > 0.05$); nevertheless, the turbidity mean WQI score has significant differences among the locations, despite that the quality interpretation for all was the same (Medium). Figure 5f shows the distribution for FC WQI scores. The mean value presents significant differences between CL and PL-CHL ($F = 4.5$, $p < 0.05$). CL presents good quality, while CHL and PL have excellent quality. The latter are principally explained by the results obtained in the Cahuita National Park.

Discussion

Each of the nine national parks or protected areas have unique hydrologic settings, making strict comparisons of WQI challenging. However, an individual comparison of each system to USEPA and WHO drinking water standards helps to guide the discussion about human health concerns and solutions (Table 1).



Figures 5 Comparison of Pacific coast (PL), Caribbean coast (CL), and Central Highlands (CHL) with the relationship of **a** overall WQI, **b** nitrates, **c** phosphates, **d** dissolved oxygen saturation, **e** turbidity, and **f** fecal coliform. The *x*-axis represents values in percentages

A major difference between the eastern and western coasts is the quality of sanitation infrastructure. Notably, a large-scale (Magnitude 7.6) earthquake event occurred in 1991 with its epicenter in Limón (Fig. 1). This event devastated the functioning sanitation systems that have not been completely, or properly, rebuilt. Additionally, due to small population densities and risk of other natural disasters on the Caribbean coast (e.g., flooding), the country has not previously invested financial capital into sanitation. Consequently, all of the sampling locations on the Caribbean coast were found to have the lowest average overall WQI for both 2018 (Tortuguero National Park at 75.0%) and 2019 (Cahuita National Park, 69.5%; Gandoca-Manzanillo, 77.6%) (Fig. 2). While Tortuguero National Park had the

lowest dissolved oxygen reported for the study, the values are likely more representative of stagnant water found in mangrove forest environments than contamination. Similarly, higher amounts of nitrates (6.0 mg/L) and phosphates (3.0 mg/L) present in mangrove environments may indicate a buildup effect. This situation can be detrimental to long-term growth rates and sustainability of such forests over time as nitrates and phosphates become immobile (Reef et al. 2010). These effects were also seen at the Pacific coast equivalent, Térreba Sierpe National Wetland. The overall WQI averages of Cahuita National Park and Gandoca-Manzanillo National Park were also greatly affected by the elevated levels of nitrates (upward of 50.0 mg/L at CH1) and phosphates (upward of 27.0 mg/L at CH2 and CH3); however, this input

Table 1 A comparison chart of USEPA and WHO drinking water standards for the parameters used in the study (USEPA 2010; WHO 2011)

Chemical parameter	EP (mg/L)	WHO (mg/L)
Nitrate (NO ₃)	0.01–3.0	0.01–5.0
Phosphate (PO ₄)	< 0.1	0.1
Physical parameter		
pH	6.5–8.5	6.5–8.0
Dissolved oxygen	4.0–9.0	6.0–8.0
Turbidity (NTU)	< 5	5
BOD	< 1.0	< 1.0
Temperature	None	None
Microbiological parameter		
<i>Escherichia coliform</i> count	< 1.0*	< 1.0*

*Units in cfu/100 mL

is probably the result of upgradient fertilizer input from large-scale banana and pineapple plantations. These values reported for the nitrates and phosphates are the highest reported of all sample locations. Additionally, fecal coliform was found at CH4 of CH6 sites in Cahuita National Park ranging from 29.0 colonies/ml to 90.0 colonies/ml. It is plausible that the presence of fecal coliforms come from nearby city sewage drainage as previously mentioned (Mena-Rivera and Quirós-Vega 2018). The only other field sites beyond Cahuita National Park that contained fecal coliform was in Corcovado National Park during the 2019 sampling season.

A drastic change in water quality was seen between the two sampling years at Monteverde Cloud Forest Reserve where the overall average WQI decreased by 14.0% from 2018 values that were ranked good. This could be due to a reduction in rainfall for the month of May 2019, where the total was 30 cm lesser than the historical monthly average. Moreover, data suggest that there have been standing impacts on the Monteverde area from the increased use of the Arenal Hydroelectric dam (Nadkarni and Wheelwright 2000; Hartshorn et al. 1982). Specifically, when energy is produced the downstream discharge is altered to the Sendero Río as it exits Lake Arenal. The Sendero Río bisects the forest reserve and acts as the main surface water source. In May 2019, dam production was at top facility production which may have had implications on stream holding capacities since rainfall was limited. Moreover, for both years, the nitrates and phosphates were elevated at this forest reserve which could also be the result of fertilizer application at adjacent coffee plantations. Naturally occurring slope instability issues within the park and the adjacent areas in 2019 may have resulted in an increase in sediment supply that contributed to higher turbidity levels.

The lake reservoir nearest to Arenal Volcano National Park was found to have high levels of turbidity and nitrates.

Although this location was sampled during the start of the rainy season, the lack of precipitation for months prior caused the lake to be at extremely low levels. Thus, energy produced at the hydroelectric facility during dry conditions resulted in the mixing of basement sedimentation. Additionally, the temperature influx from geothermal venting from Volcano Arenal may stimulate the growth of nitrate producing microbial life (Sand 2004).

The acidic waters found at Braulio Carrillo National Park result from the highly felsic volcanic ash deposits found in the area from the nearby Barva Volcano that is currently dormant. Likewise, the fine-grained ash gets suspended in the rivers that cross-cut the park and can influence the water clarity.

Similar to Arenal Volcano National Park and Braulio Carrillo National Park, Tenorio Volcano National Park is also influenced by recent volcanic activity. In particular, the park has active fumaroles releasing gaseous sulfur dioxide to the surface which speaks of the acidic waters at some sites. The decrease in water clarity may be a result of silica precipitation that makes the water opaque in Río Celeste.

The water sampled at Térraba Sierpe National Wetland had the highest acidity of any of the nine sample locations. Despite the mangrove forest having the ability to trap calcium-rich shells and offshore coral to make brackish waters alkaline, mangrove soils tend to be slightly acidic due to the presence of sulphur-reducing bacteria that form on their extensive root systems. Likewise, this location also has the lowest recorded dissolved oxygen levels comparatively. Low dissolved oxygen is common in mangrove forests as water becomes trapped to create anoxic zones. Moreover, oxygen found in between the particles of the saturated soil are used up by the decay and respiration of bacteria. The oxygen content is only replenished by the circulation of tidal water and exchange with the atmosphere (Goodwin et al. 2001). Nitrates, although needed in the sustainability of a mangrove forest, are typically found at low levels, as denitrifying bacteria are abundant in mangrove soils. Elevated levels at Térraba Sierpe National Wetland may be the result of major deforestation (loss exceeding 6333 acres) through the disturbance of the nutrient cycle process. Bormann and Likens (1994) explain that land-clearing causes an oversupply of nitrate in the upper soil layers; nitrates that were once stored were then able to be added to the Río Grande de Térraba from surface water runoff. Excess nitrates are of concern, as nutrient enrichment is a major threat to marine ecosystems as it limits the ability for mangroves to grow (Sarker et al. 2019; Reel et al. 2010).

Corcovado National Park is a remote sampling destination on the Osa Peninsula along the Pacific Ocean. Two different sampling locations were visited during the 2018 and 2019 field seasons. The lower value of overall WQI in 2018 (81.0%) compared to 2019 (84.2%) is the result of higher

levels of nitrates, phosphates and turbidity. However, these elevated levels are likely to be the result of natural decay processes releasing nitrates, mineralization of compounds in soils and sediments to produce traceable levels of phosphates, quick moving streams that heighten erosion resulting in turbid water conditions and local gray water disposal. Although the overall WQI for 2019 was good, the Sirena Biological Field Station has increasingly become a tourist destination for school groups, whereby increasing the need for waste disposal. Poorly constructed outhouses were built adjacent to the Park's major river (e.g., Río Claro) to meet this demand, but it has led to fecal contamination as seen at C5B, C6B, and C7B.

Conclusions

The aim of this project was to spearhead forest conservation that is supported by scientific substantiation. This research supports the planning and development of Costa Rica water resources, whereby linking management and sustainable use described by Guzman-Arias and Calvo-Alvarado (2013). Additionally, as the authors have witnessed from their research in emerging countries, it is possible to monitor environmental changes with the application of water quality assessments to gauge system degradation (Jovanelly et al. 2012, 2015).

From this study, we show that the overall health of the surface waters in nine national parks or protected areas are in good standing. In most scenarios, the elevated levels of nitrates, phosphates, or turbidity, or the depleted levels of dissolved oxygen, were the result of natural processes, and not human disturbance. However, in parks that are heavily influenced by dam release (i.e., Monteverde Cloud Forest Reserve), agriculture (i.e., Cahuita National Park) or tourism (i.e., Corcovado National Park, specifically Sirena Biological Station), our study has shown that the well-being of the watersheds is deteriorating by these system impacts. Additionally, climatic variability, especially changes in precipitation patterns, critically influence water quality. For this reason, the authors feel that it is important to establish a monitoring network among these watersheds to have long-term data within the different seasons that will provide a better understanding of the season variability. At present, there are several investigations in urban watersheds (e.g., Mena-River et al. 2017) that show lower levels of WQI in dry season; however, current investigation in non-urban catchments show lower values in the wet season as mainly related by the agricultural runoff.

A major conclusion drawn from this study is that that baseline water quality data are critical for evaluating system integrity and can act as an early indicator to challenges that a forest may face. (An example garnered from this study

would be the extremely elevated levels of nitrogen in the Térraba Sierpe National Wetland that could potential cause the ecosystem to perish.) Therein, it needs to be recognized that aquatic systems ultimately support the biodiversity of a forest ecosystem (Hjältén et al. 2016). Stressors such as stream flow regulation, stream channelization, pollution, and climate change affect the water and will translate into diminished ecosystem response.

Whilst Costa Rica is a world leader in the use of carbon neutral energy, biodiversity, and environmental protection, the need for baseline water quality data collection as part of a forest assessment cannot be overlooked and should be deemed compulsory for sustainability and ecosystem security. Physical and chemical data collected on surface water (as done in this study) could be used as a tool to monitor for system change. Moreover, the further use of the WQI to communicate science without jargon in Costa Rica can protect biodiversity by promoting a multi-faceted grass-roots approach to conservation that is community centered, whereby local control of monitoring reserves and managing databases motivates pro-environmental behavior and policy. Studies have shown that increased environmental knowledge often leads to civic action in environmental/conservation efforts (Hungerford 2010; Johnson-Pynn and Johnson 2010; Mugisha 2002). Directly relating to Costa Rica, this idea of shared scientific evidence could be used to promote accountability of the ecotourism companies to work toward best practices as suggested by the World Bank (2019) and Garen (2000). Together, these components support long-term, sustainable benefits to people and the environment.

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Conflict of interest All authors declare that they have no conflict of interest.

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